

BROOKHAVEN NATIONAL LABORATORY
MEMORANDUM

DATE: July 17, 1980

TO: Hugh Pratt
FROM: E. T. Lessard *ETL*
SUBJECT: Chronic Dosimetry

We have enclosed a recently assembled draft of "Reconstruction of Chronic Dosimetry for Residents of Rongelap and Utirik Atolls" for your consideration. The information in this report makes use of data collected by yourself and members of your medical team. Your comments would be greatly appreciated.

During the next few weeks, members of the MIRSP team will be travelling to the Marshall Islands, and to Seattle, Washington to the Annual Health Physics Society Meeting. Thus, your timely review will afford us the opportunity to incorporate changes in the final draft.

Thank you in advance for any suggestions you make.

ETL:js
Encl:

cc: N. A. Greenhouse
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^{137}Cs . Four persons visited Argonne from Rongelap and, in addition, pooled urine samples from both atolls were analyzed radiochemically for ^{137}Cs and ^{90}Sr . Subsequent Brookhaven National Laboratory expeditions by the Medical Department and Safety and Environmental Protection Division utilized whole body counting and radiochemical analysis of urine and blood samples to identify and quantify the radionuclides that were present in the body. The results of these radiological measurements are given in terms of body burden in tables 1 and 2.

The aforementioned body burden tables illustrate adult mean values for Rongelap and Utirik. An adult, as classified here, was a person over 16 years of age. The mean body mass in this age interval was 60 kilograms. The observed body mass versus age distribution is shown in figure 1 for Rongelap residents. The same body mass versus age distribution was observed at Utirik.

Due to the paucity of measurements at Utirik, information on ^{60}Co , ^{65}Zn and ^{55}Fe was in some instances derived from the ratio of adult mean body burdens between Rongelap and Utirik. A mean ratio of 2.6 was observed in body burdens for ^{65}Zn , ^{90}Sr and ^{137}Cs after they reached their maximum values. The standard deviation of this ratio was 15%.

In the following analysis, personal body burden histories and residence intervals, in conjunction with contemporary dosimetric models, are used to estimate internal dose. Dosimetric distributions were constructed from the results and a summary of the derived activity ingestion rates and dose equivalents was provided for various subgroups of the population. Additionally, exposure rate history curves were constructed for each atoll for the period following the BRAVO test. These data, together with appropriate conversion factors and living pattern models, provided an estimate of external dose equivalent.

Table 1

Rongelap Body Burdens

	Adult Males		Adult Females		Adults		Days Post Return Days
	Body Burden μCi	Number of Persons	Body Burden μCi	Number of Persons	Body Burden μCi	Number of Persons	
^{60}Co	2.9×10^{-5}	NA	1.7×10^{-5}	NA	2.3×10^{-5}	NA	1
	1.0×10^{-2}	37	7.8×10^{-3}	37	9.0×10^{-3}	74	1370
	2.5×10^{-3}	45	2.0×10^{-3}	45	2.2×10^{-3}	90	2831
^{65}Zn	4.3×10^{-2}	NA	3.8×10^{-2}	NA	4.1×10^{-2}	NA	1
	4.3×10^{-1}	30	3.8×10^{-1}	12	4.1×10^{-1}	42	304
	6.2×10^{-1}	32	5.0×10^{-1}	27	5.6×10^{-1}	59	639
	9.5×10^{-2}	38	8.5×10^{-2}	23	9.0×10^{-2}	61	1370
^{55}Fe	4.3×10^{-1}	28	4.0×10^{-1}	32	4.1×10^{-1}	60	4626
^{90}Sr	1.9×10^{-4}	NA	1.4×10^{-4}	NA	1.7×10^{-4}	NA	1
	3.7×10^{-3}	11	2.8×10^{-3}	4	3.4×10^{-3}	15	304
	5.7×10^{-3}	24	3.5×10^{-3}	16	4.8×10^{-3}	40	639
	3.7×10^{-3}	9	1.6×10^{-3}	4	3.0×10^{-3}	13	1370
	8.8×10^{-3}	12	7.9×10^{-3}	13	8.4×10^{-3}	25	2100
	7.9×10^{-3}	11	7.4×10^{-3}	7	7.7×10^{-3}	18	2466
	2.8×10^{-3}	12	4.6×10^{-3}	12	3.7×10^{-3}	24	3561
	3.9×10^{-3}	11	3.1×10^{-3}	11	3.5×10^{-3}	22	3927
	4.1×10^{-3}	11	3.3×10^{-3}	13	3.6×10^{-3}	24	4292
	1.3×10^{-3}	8	3.3×10^{-3}	11	2.5×10^{-3}	19	4657
	3.1×10^{-3}	8	2.8×10^{-3}	7	3.0×10^{-3}	15	5022
	2.0×10^{-3}	5	1.4×10^{-3}	7	1.6×10^{-3}	12	5388
	6.6×10^{-3}	4	4.2×10^{-3}	7	4.3×10^{-3}	13	5753
	3.3×10^{-3}	10	1.7×10^{-3}	4	2.8×10^{-3}	14	6118
	4.4×10^{-3}	23	NA	0	NA	NA	7579
	0.3×10^{-4}	24	4.6×10^{-4}	19	5.5×10^{-4}	43	8097
^{137}Cs	1.4×10^{-2}	NA	8.4×10^{-3}	NA	1.1×10^{-2}	NA	1
	8.7×10^{-1}	NA	5.2×10^{-1}	NA	6.8×10^{-1}	NA	304
	7.9×10^{-1}	47	4.1×10^{-1}	49	5.7×10^{-1}	96	639
	9.5×10^{-1}	37	4.7×10^{-1}	37	6.7×10^{-1}	74	1370
	9.4×10^{-1}	44	4.9×10^{-1}	45	6.8×10^{-1}	89	2831
	4.8×10^{-1}	22	3.0×10^{-1}	24	3.9×10^{-1}	46	6118
	3.0×10^{-1}	30	1.9×10^{-1}	21	2.5×10^{-1}	51	7213
	1.8×10^{-1}	19	1.5×10^{-1}	18	1.7×10^{-1}	37	8097

NA = Not analyzed

Table 2

Utirik Body Burdens

	<u>Adult Males</u>		<u>Adult Females</u>		<u>Adults</u>		<u>Days Post Return Days</u>
	<u>Body Burden μCi</u>	<u>Number of Persons</u>	<u>Body Burden μCi</u>	<u>Number of Persons</u>	<u>Body Burden μCi</u>	<u>Number of Persons</u>	
^{60}Co							
D	4.0×10^{-3}		3.1×10^{-3}		3.5×10^{-3}		2464
D	9.7×10^{-4}		7.6×10^{-4}		8.7×10^{-4}		3924
^{65}Zn							
	3.5×10^{-1} *	2	-		-		
	2.7×10^{-1}	14	1.6×10^{-1}	15	2.1×10^{-1}	29	1734
D	3.7×10^{-2}		3.3×10^{-2}		3.5×10^{-2}		2464
^{55}Fe							
D	1.7×10^{-1}		1.6×10^{-1}		1.6×10^{-1}		6114
^{90}Sr							
	1.4×10^{-3}	5	2.4×10^{-3}	2	1.7×10^{-3}	7	1734
	1.2×10^{-3}	5	1.3×10^{-3}	6	1.3×10^{-3}	11	7213
	NA	12	NA	12	NA	24	8669
	1.5×10^{-4}	14	1.5×10^{-4}	17	1.5×10^{-4}	31	9225
^{137}Cs							
	4.1×10^{-1}	NA	2.7×10^{-1}	NA	3.3×10^{-1}	NA	1004
	2.9×10^{-1}	15	2.0×10^{-1}	15	2.5×10^{-1}	30	1734
	2.6×10^{-1}	9	1.3×10^{-1}	13	1.8×10^{-1}	22	7213
	1.2×10^{-1}	27	7.8×10^{-2}	21	1.0×10^{-1}	48	8309
	6.2×10^{-2}	19	4.3×10^{-2}	17	5.3×10^{-2}	36	9225

D = Ratio derived body burden

NA = Not analyzed

* = Measured at Argonne National Laboratory

METHODS

DOSIMETRIC EQUATIONS

Declining continuous uptake of radioactive dietary items was mathematically modeled for each nuclide of concern. The following general equations were used

$$\lambda P^{\circ} = \frac{U U_s / f_u - q^{\circ} \left(\sum_i K_i \chi_i' e^{-(\lambda+K_i)t} \right)}{f_l \left(\sum_i \frac{\chi_i K_i}{K_i - K_E} \left(e^{-(\lambda+K_E)t} - e^{-(\lambda+K_i)t} \right) \right)}, \text{ or} \quad (1)$$

$$\lambda P^{\circ} = \frac{q - q^{\circ} \left(\sum_i \chi_i' e^{-(\lambda+K_i)t} \right)}{f_l \left(\sum_i \frac{\chi_i}{K_i - K_E} \left(e^{-(\lambda+K_E)t} - e^{-(\lambda+K_i)t} \right) \right)}, \text{ and} \quad (2)$$

$$D = f_l \lambda P^{\circ} \sum_i \frac{\chi_i}{K_i - K_E} \left(\frac{K_i - K_E - (\lambda + K_i) e^{-(\lambda+K_E)t} + (\lambda + K_E) e^{-(K_i + \lambda)t}}{(K_E + \lambda)(K_i + \lambda)} \right) + q^{\circ} \sum_i \frac{\chi_i'}{\lambda + K_i} \left(1 - e^{-(\lambda+K_i)t} \right), \quad (3)$$

where

$t \equiv$ time post onset of uptake, days,

$\lambda \equiv$ instantaneous fraction of atoms decaying per unit time, day^{-1}

$P^{\circ} \equiv$ initial atom ingestion rate, atoms day^{-1} ,

$K_i \equiv$ instantaneous fraction of atoms removed from compartment i by physiological mechanisms, day^{-1} ,

$\chi_i \equiv$ compartment i deposition fraction,

$\chi_i' \equiv$ the number of atoms in compartment i relative to the number in all compartments at the onset of uptake ($t=0$),

$U \equiv$ instantaneous urine activity concentration, $B_q \ell^{-1}$,

$U_s \equiv$ subject urine excretion rate, $\ell \text{ day}^{-1}$,

$f_l \equiv$ fraction from GI tract to blood,
 $f_u \equiv$ fraction excreted by the urine pathway,
 $K_E \equiv$ instantaneous fraction of atoms removed or added to the atom uptake per unit time, day^{-1} ,
 $q \equiv$ instantaneous body burden, Bq,
 $q^0 \equiv$ body burden at the onset of uptake, Bq,
 $D \equiv$ the number of disintegrations in all compartments occurring during the uptake interval, Bq days.

Equations 1 and 2 were used to determine the dietary removal rate constant K_E and then the initial daily activity ingestion rate required to produce the measured or derived body burden. Equation 3 was used to determine the number of disintegrations that occurred in the body during the residence interval of an individual living on Rongelap or Utirik Atoll.

DETERMINATION OF K_E

If the mean residence time in the diet is much much longer than the residence interval, then constant continuous uptake is achieved. Equations 1 and 2 can be converted to the constant continuous equations by replacing K_E with $-\lambda$. Single uptake expressions are obtained by setting P^0 equal to zero. In some cases only radioactive decay may remove the nuclide from dietary items; for these cases K_E would equal zero. In the case of the former Bikini residents, the maturing of coconut trees during residence on Bikini Atoll caused a continuously increasing dietary uptake of ^{137}Cs . Thus, K_E was found to have a negative value. In the case of Rongelap and Utirik, K_E was found to have a positive value for ^{137}Cs , ^{65}Zn , ^{60}Co and ^{90}Sr . This indicated that in addition to radioactive decay, some other removal mechanism decreased the radioactivity in

dietary items during the residence interval. For the nuclide ^{55}Fe , only one measurement in blood was published by the BNL Medical Program (Be72); thus an estimate of K_E was not possible.

K_E was determined by using equation 1 or 2 and the population subgroup mean body burdens or urine activity concentrations. A portion of these bioass. data are illustrated for adult males and females in Figures 2, 3, 4, 5 and 6. Two consecutive urine or body burden data points were used to eliminate the unknown ingestion rate from the equation. This method yields $n-1$ estimates of K_E where n was the number of data points. An average value of K_E was assigned for each nuclide and the results for the Rongelap and Utirik populations are given in Table 3. For the evaluation of K_E from equations 1 and 2, radiological and physiological parameters were obtained from the open literature (ICRP59, ICRP6 ICRP69, ICRP79, Ki78). A representative sample of these parameters is presented in Table 4.

Table 3

	Summary of Dietary Rate Constants (K_E, d^{-1})			
	^{60}Co	^{90}Sr	^{65}Zn	^{137}Cs
Rongelap Adults				
Males	1.5×10^{-3}	1.8×10^{-4}	3.1×10^{-3}	1.4×10^{-4}
Females	1.6×10^{-3}	4.1×10^{-4}	3.5×10^{-3}	1.4×10^{-4}
Adults	1.5×10^{-3}	1.9×10^{-4}	3.1×10^{-3}	1.4×10^{-4}
Utirik Adults				
Males	N.D.	4.6×10^{-4}	N.D.	1.4×10^{-4}
Females	N.D.	4.0×10^{-4}	N.D.	1.4×10^{-4}
Adults	N.D.	4.2×10^{-4}	N.D.	1.4×10^{-4}

N.D. = No data sufficient for analysis.

Table 4

Total Body Dosimetric and Physiologic Data

Nuclide	Compartment Deposition Fraction	Compartment Removal Rate Constant	GI Tract to Blood Transfer	Fraction Excreted in Urine	Decay Constant	Significant Progeny	Branching Ratio
A_ZX	χ_i	$K_{i,d^{-1}}$	f_l	f_u	$\lambda_{d^{-1}}$	A_ZX	
${}^{137}_{55}\text{Cs}$	0.13 0.87	0.50 0.0051	1.0	0.90	6.3×10^{-5}	${}^{137m}_{56}\text{Ba}$	0.946
${}^{65}_{30}\text{Zn}$	0.25 0.75	0.058 0.0022	0.35	0.25	2.8×10^{-3}	${}^{65*}_{29}\text{Cu}$	0.49
${}^{90}_{38}\text{Sr}$	0.89 0.059 0.051	0.21 7.1×10^{-4} 1.0×10^{-4}	0.20	0.85	6.5×10^{-5}	${}^{90}_{39}\text{Y}$ ${}^{90*}_{40}\text{Zr}$	1.0 0.0002
${}^{60}_{27}\text{Co}$	0.5 0.3 0.1 0.1	1.4 0.12 0.012 8.7×10^{-4}	0.05	0.70	3.6×10^{-4}	${}^{60*}_{28}\text{Ni}$	1.0
${}^{55}_{26}\text{Fe}$	1.0	3.5×10^{-4}	0.1	0.0	7.0×10^{-3}		

The values of K_E were similar for males and females and for residents of Rongelap and Utirik. For ^{90}Sr on Rongelap a factor of 2 difference between K_E values was observed for males and females. The female parameter for Rongelap Atoll compares with that obtained from the Utirik data. A paired t-test of the Rongelap male and female data indicates that the male/female difference was highly probable and therefore not significant. This difference leads to a bimodal activity ingestion rate distribution for ^{90}Sr in the Rongelap population.

Data for ^{60}Co and ^{65}Zn was not sufficient for analysis for the Utirik Atoll residents. Values for K_E observed at Rongelap were assigned to Utirik males and females and body burden histories for population subgroups were reconstructed using equation 1 or 2. Figures 7 and 8 illustrate the derived mean adult body burdens for all significant nuclides studied on Rongelap and Utirik. This method provides a best fit of the data shown in figures 2 through 6, and provides a body burden history during the early years post return at Utirik, a time when body burden measurements were not made. Actual data points are also plotted to demonstrate the fit.

The curves shown for ^{55}Fe in figures 7 and 8 were obtained by setting K_E equal to zero. This underestimated the initial body burdens and overestimated future ones. Since ^{55}Fe contributed less than 1.0% to the total dose equivalent an arbitrary assignment of K_E based on observed values for the other nuclides was not attempted. During 1974, another series of blood samples was obtained from Rongelap and Utirik (Co75). Analysis for ^{55}Fe was to be performed; however, no records have been located. Once these records are found, a reanalysis of ^{55}Fe and its impact on early dose equivalent rates will be conducted. A substantial change in dose equivalent is not to be expected.

Figure 4 and figure 6 illustrate the observed adult histories of ^{90}Sr and ^{137}Cs mean urine activity concentrations. Mean values for adult males or all adults were plotted. Measured values for ^{137}Cs body burdens were also shown in figure 7. A much smoother curve was plotted in figure 7 and it was determined that the collection and analysis technique for urine samples introduced the additional variations. Based on this observation for ^{137}Cs , a smooth body burden curve for ^{90}Sr , reconstructed from raw data and equation 1, was considered a more accurate history. A detailed presentation of the greater variation in urine bioassay measurements versus direct body burden measurements can be found in Mi81.

Figure 9 illustrates the variation exhibited in the body burden of 5 randomly chosen subjects over the 25 year monitoring period. These individual variations may have had a dramatic impact on the mean data. In figure 2, which illustrates the adult male, adult female and adult population mean ^{137}Cs body burden for the 25 year exposure period, a decrease followed by an increase was seen during the years 1958 through 1963. Although the Castle BRAVO test initially contaminated Rongelap, it had been proposed that the Hardtack Phase I series added to this an amount of contamination equal to that responsible for the figure 2 body burden pattern (Co63). Figure 9 suggests that most individuals counted in those years had maintained or declined in body burden; however, one individual's burden (#881 M) rose and fell quite differently from the others. Several factors could have contributed to this variation from the mean such as departure and return to the atoll, sickness, the dietary contribution of imported foods, etc.. Since the mean values are based on small numbers of persons who were chosen at random, it is conceivable that individuals like 881 M

influenced the mean body burdens to a greater degree than recontamination of the inhabited atolls. This conclusion is moot since the body burden of an individual was not monitored consistently throughout the residence interval except in the few cases exhibited in figure 9.

RESULTS AND DISCUSSION

DAILY ACTIVITY INGESTION RATES

Daily activity ingestion rates were calculated for dosimetrically significant nuclides post return. An exponential decline was proposed for the ingestion rate within a population subgroup and initial reference values are given in figures 10, 11, 12, 13 and 14 (June 1, 1957 was assigned as a return date to Rongelap). Figure 10 demonstrates the difference in ingestion of ^{137}Cs for various population subgroups. This undulating pattern was exhibited by ^{137}Cs , ^{90}Sr and ^{65}Zn , nuclides for which sufficient data existed for analysis.

Differences in ingestion rates of the stable element at the same geographic location have been shown to occur among members of a population (ICRP 23). Age dependent diet studies for ingestion of Cs for urban Japan have values varying from $11 \mu\text{g d}^{-1}$ for adults to $8.6 \mu\text{g d}^{-1}$ for children. Sr in a western-type diet rose from $600 \mu\text{g d}^{-1}$ for infants to $690 \mu\text{g d}^{-1}$ for 5 year olds to $3,600 \mu\text{g d}^{-1}$ for 13 year olds and fell to a mean of $1,900 \mu\text{g d}^{-1}$ for adults. Zn in the United Kingdom rose from 2 to 40mg d^{-1} , the higher value of Zn being observed in adult tea drinkers. Fe ingestion in a western-type diet has a minimum at age 3 and maximums at ages 1 and 20 years. Co is ingested at a rate of $20 \mu\text{g d}^{-1}$ for Japanese adults and half this amount for children. The Marshallese population also exhibits dietary changes as a function of age. The authors of the Marshall Islands Diet and Living Pattern Study (Na80) observed coconut sap being used as a major food supplement for infants, and later in adult life as a major source of daily fluid intake. Since coconuts and coconut tree sap provided the major source of ^{137}Cs on Bikini Atoll (Le80, Mi80), the shape of Figure 10 was in agreement with the observed diet pattern.

Figure 11 compiled the individual data calculated for ^{137}Cs for all Rongelap residents and was referenced to June 1, 1957. The individual maximum ^{137}Cs daily activity ingestion rate was approximately 4 times the population mean value. The standard deviation observed for the adult activity ingestion rate distribution was 41% of the mean value, 39% of the mean value for young adults, 48% for adolescents, 38% for children and 54% for infants. Adolescents and infants exhibited a broader distribution than adults while children showed a fractional variation in activity ingestion rate similar to adults. Breast feeding versus coconut sap supplements would have contributed to the greater variation observed in infants. Adolescents and young adults were the population subgroups which have been observed to move frequently between atolls. This mobility would lead to greater variations in the daily activity ingestion rates relative to those observed in the more stationary population subgroups.

Figure 12 also exhibited a wave pattern; however, a distinct difference between males and females was indicated. This difference arose from the use of dietary rate constants listed in Table 3 which were derived from urine data for male and female residents at Rongelap Atoll. Its major impact was on the dose equivalent rate, not on the total dose equivalent; and its effect was to cause the dose equivalent rate for males to rise and decline more rapidly than for females.

Figures 13a and 136 summarize the individual data for ^{90}Sr for all Rongelap residents and were referenced to June 1, 1957. A bimodal shape was observed for the distributions which contained both sexes thus reflecting the difference in the ^{90}Sr dietary rate constants. Data from urine bioassay indicated that the observed difference between the male and female value for K_E was not significant. A t-test was performed between consecutive urine measure-

ment data during the twenty-three year residence interval. The results indicate that because of urine activity concentration variability, a probability of 6 out of 10 existed such that the male value for K_E would be different from the female value by the factor observed. Thus differences in the derived activity ingestion rates and dose equivalents were not significant.

Figure 14 shows a semi-log plot of the ^{65}Zn and ^{137}Cs activity ingestion rate histories for adults on Rongelap. A smooth curve was drawn between points and the appearance of an increasing ^{137}Cs ingestion rate during the 1960's indicates the possibility of another contaminating event. The Hardtack Phase I series was conducted just prior to the observed increase in the curve and fallout from the Cactus, Yellow Wood and Hickory experiments detonated at Bikini and Enewetak would have reached Rongelap. However, several observations fail to support the conclusion that recontamination was significant. These are as follows: 1) the increase in ^{137}Cs ingestion rate was not in conjunction with an increase of ^{65}Zn ; however, since ^{65}Zn is an activation product it may have not been produced in the same proportions 2) the peak ^{137}Cs body burden at Utirik occurred nearly three years after the initiating event, Castle BRAVO, while the peak body burden at Rongelap followed six years after the potentially contaminating experiments of the Hardtack series in 1958 3) the activity ingestion rate at Utirik demonstrated a continuously declining pattern versus the humped pattern observed at Rongelap. This occurred even though there was an equal external exposure rate history following the Hardtack series as measured by the U.S. Public Health Service on both Rongelap and Utirik (Un 59). These facts suggest that the Hardtack series was not a major factor influencing the Rongelap body burden patterns. Thus it is assumed that persons who had body burdens significantly different from the mean body burden for the population

caused the extent of variation reported. Based on these observations a smooth description of the body burden and activity ingestion rate was adopted and a declining continuous uptake model was used to generate the curves in figures 7 and 8.

INTERNAL DOSE EQUIVALENT RATES

The approximate instantaneous dose equivalent rates for the total body were determined from the body burden data illustrated in figures 7 and 8 and from the following equation

$$\dot{H} = q I, \quad (4)$$

where

$\dot{H} \equiv$ the total body dose equivalent rate, mRem y^{-1} ,

$I \equiv$ equilibrium dose equivalent rate to the total body per unit body burden, mRem $y^{-1} \mu Ci^{-1}$,

$q \equiv$ instantaneous body burden, μCi .

The approximate nature of the estimate was due to the assumption that the radioactive atoms were distributed among the body tissues as they would be following constant continuous uptake for periods of time much greater than the mean residence time for the total body. In the case of ^{90}Sr , 86% of equilibrium was assumed. These assumptions were not used in the estimate of the total dose equivalent. In addition, since mean adult body burdens were computed, a factor of 1.2 was needed to adjust for differences in body mass relative to a 70 kg adult. Table 5 lists values of I which were determined from information given in ICRP59 and corrected for body mass differences.

Figure 15 illustrates the relative contribution to the composite dose equivalent rate for each dosimetrically significant internally deposited nuclide. For the average Rongelap adult, the residence interval begins June 1, 1957; however, many adults were reported to have resettled during the next 3-6 months (Co80b). The composite dose equivalent rate indicated that a broad

Table 5

$\begin{matrix} A \\ Z \\ X \end{matrix}$	Total Body Equilibrium Dose Equivalent Rate per Unit Total Body Burden,
	mRem y ⁻¹ μCi ⁻¹
$\begin{matrix} 55 \\ 26 \\ \text{Fe} \end{matrix}$	2×10^0
$\begin{matrix} 60 \\ 27 \\ \text{Co} \end{matrix}$	6×10^2
$\begin{matrix} 65 \\ 30 \\ \text{Zn} \end{matrix}$	1×10^2
$\begin{matrix} 90 \\ 38 \\ \text{Sr} \end{matrix}$	3×10^2
$\begin{matrix} 137 \\ 55 \\ \text{Cs} \end{matrix}$	2×10^2

maximum of approximately several hundred millirem per year persisted for several hundred days. The majority of the dose rate is attributable to the ^{137}Cs component. Cesium dominated over the entire post return period and would be of prime concern for populations returning to a contaminated environment years after a fission-type initiating event.

Figure 16 illustrates two possibilities for the Utirik body burden history during the first three years post return. The higher body burden resulted from use of the two measured ^{65}Zn body burden means for adults on Utirik and the observed dietary rate constant from Rongelap. It was observed on Rongelap that .031% of ^{65}Zn was removed from the diet pathway each day in addition to radioactive decay. Additionally, reduction in dietary radioactivity on Rongelap had been observed for ^{137}Cs , ^{90}Sr and ^{60}Co to be greater than that predicted by radioactive decay alone. Instantaneous reduction fractions very similar to

Rongelap were observed at Utirik for the ^{90}Sr and ^{137}Cs nuclides. The lower curve on Figure 16 reflects the dose equivalent, dose equivalent rate and body burden which would have occurred had radioactive decay alone accounted for the removal of ^{65}Zn from the Utirik environment. Since additional mechanisms could be measured for other nuclides at Utirik and for the ^{65}Zn nuclide on a nearby atoll, the upper curve was chosen as the most likely body burden history for adults post return to Utirik Atoll.

Figure 17 indicates the Utirik adult mean total body dose equivalent rate for each nuclide. An obvious difference relative to the Rongelap history exists, ^{65}Zn not ^{137}Cs was the major nuclide contributing to the dose equivalent rate. This was due to the Utirik population returning 4 months after the initial contaminating event, and the Rongelap population returning after 3 years. The age of the fallout had a dramatic influence on the importance of each nuclide contributing to the internal dose equivalent. In fact ^{60}Co and ^{65}Zn played major roles during the first 3 years, a time interval that corresponded to the period during which field whole body counting facilities were being developed at Brookhaven National Laboratory and when medical examinations for people on Utirik Atoll were not done. Additionally, pooled and/or individual radiochemical analysis of urine was not performed during this period. The impact of ^{65}Zn and ^{60}Co was such that even if the least conservative dietary rate constant ($K_E=0$) was used for Zn, the dose equivalent rate for the average adult was in excess of Federal Radiation Council Guidelines for the first 2 years following the return to Utirik.

INTERNAL DOSE EQUIVALENTS

Disintegrations occurring in the total body of an individual during residence following repatriation were determined by several methods. Equation 3, together with personal body burden histories and atoll-specific dietary rate constants from table 3, provided an initial estimate of disintegrations between consecutive body burden measurements. The second method used was a log-log plot of the subject's body burden history and an algebraic determination of area between two consecutive measured points. The third method used a linear plot of the subject's body burden history. The area under the curve was cut and weighed and compared to a standard weight of known area. Quality control procedures required that all three methods agree within $\pm 10\%$ before a subject was assigned his or her total body disintegrations during residence post return. In general, the methods compared to within $\pm 5\%$.

After the total number of disintegrations occurring in a subjects body were assigned, they were apportioned among the body organs according to the following equation

$$F = \frac{f_2' \sum_i A_i B_i (\sum_i C_i D_i + \ln 2/\lambda)}{\sum_i C_i D_i (\sum_i A_i B_i + \ln 2/\lambda)} \quad (5)$$

where

F \equiv the fraction of total body disintegrations occurring in the organ of interest,

A_i \equiv organ compartment deposition fraction for the element,

B_i \equiv organ compartment biological half time for the element,

C_i \equiv total body compartment deposition fraction for the element,
 D_i \equiv total body compartment biological half time for the element,
 f_2' \equiv fraction of the element from blood to organ of reference.

Equation 5 applied where significant decay occurred at the deposition site, and not during transit or re-transit to the organ of interest. Values for compartment deposition fractions and compartment half times were obtained from Ki78. Values for the remaining quantities were from ICRP59.

The dose equivalents to a specific organ or the total body were determined by using the source to target dose equivalent per unit cumulated activity parameters from Ki78. The total target dose equivalent was obtained by summation of the dosimetric contributions from all source organs. Several important modifications to the general procedure were made in order to compute individual dosimetric results. For each person, the source-to-target dose equivalent per unit cumulated activity was weighted by the ratio of a standard man's body mass relative to the actual mean body mass during the interval for which the dose equivalent was determined. In the case of ^{137}Cs , the long term biological removal rate constant for the Marshallese population was highly dependent upon body mass (Mi81). Appropriate modifications to equations 2, 3 and 5 were made to reflect this dependency. Finally, for ^{90}Sr deposition in bone, 28% of the source to target dose equivalent per unit cumulated activity was assumed from cancellous bone and 72% from cortical bone.

Figure 18 demonstrates the mean dose equivalent from ^{137}Cs for various age and sex groupings. The residence interval was from 1957 to 1980 for this population. The adolescents and persons above 50 years of age in 1957 maintained the lowest dose equivalent. Persons who died during this period were not included

in the figure nor were they included in any dosimetric distributions for any of the nuclides. Thus all persons considered, regardless of initial age in 1957, experienced a 23 year exposure interval.

Figure 19 shows dose equivalent distributions according to age and sex for ^{137}Cs among the Rongelapese. The shape of the population distribution was Poisson with a mean of 1.7 Rem and a maximum of 9.0 Rem. Thus the maximum was 5.3 times the mean value for ^{137}Cs on Rongelap. An examination of the subgroup distributions reveals that persons who were infants at the time of rehabilitation at Rongelap also were the recipients of the higher doses. This was due to the combined effects of lower average body mass, a higher average ingestion rate and more rapid turnover of ^{137}Cs than that for adults or even children. The parameter having the greatest impact on the infant dose equivalent was body mass. The standard deviation for the adult males distribution was 49% of the mean dose equivalent, for adult females 43% of the mean dose equivalent, and for adolescents 47%. Within a subgroup, the maximum observed dose equivalent was approximately twice the mean value for all distributions considered here.

Figure 20 evinces mean dose equivalents as a function of returning age groups for ^{65}Zn on Rongelap. Adolescents, young adults and adults 50 and up were the groups receiving lower total dose equivalents, while children and middle aged persons received higher dose equivalents during the residence interval. Measured ^{65}Zn data for persons who were infants at the return date were not reported in the publications by Conard, et al.

Figure 21 shows the dosimetric distributions observed for members of the Rongelap population for ^{65}Zn . Again the population overall exhibited a Poisson distribution of dose with a maximum value nearly three times the mean. Children demonstrated higher doses than persons who were adults during the entire 23 year

period. The standard deviation was in general 30% of the mean value for all age and sex subgroup distributions. This less pronounced variation may be due to the fact that ^{65}Zn measurements took place over a three year interval while ^{90}Sr and ^{137}Cs occurred over a 23 year interval and thus was contained in a more homogeneous population than were the longer lived nuclides.

Figures 22 and 23a and 23b summarize the ^{90}Sr dose equivalent results for individuals at Rongelap.

In this analysis only the ingestion pathway was considered important. Although some radioactivity would enter the body via resuspension and direct inhalation pathways, it is known that the ratios of food and fluid intake to blood relative to airborne intake to blood, for the stable naturally occurring analogs to the radionuclides considered here, are as follows:

Co > 3000	Zn > 130
Fe 550	Sr > 10,000
Cs 400	

EXTERNAL EXPOSURE

A value of .73 Rads in tissue of interest per Roentgen measured in air at one meter above the surface was used to convert exposure in air to absorbed dose in tissue. The source was assumed to be an exponential distribution of ^{137}Cs activity with depth in soil, typical of aged fallout (Be70). Because of the multidirectional nature of the source, variation of absorbed dose with depth of organ was minimal. Additionally, external doses were adjusted for living pattern variations since the atolls present a heterogeneous exposure rate environment (Gr77).

External exposure calculations are based on figures 24, 25 and 26 which were derived from data listed in Cr56, Sh57, Un59, and Gr77. The area under straight line portions of the curve was determined by

$$X = \frac{R_2 t_2 - R_1 t_1}{n + 1}, \quad (6)$$

where

$X \equiv$ external exposure during straight line interval, mR,

$R_2 \equiv$ exposure rate at the end of the interval, mRh^{-1} ,

$R_1 \equiv$ exposure rate at the beginning of the interval, mRh^{-1} ,

$t_2 \equiv$ time post detonation at the end of interval, hours,

$t_1 \equiv$ time post detonation at the beginning of interval, hours,

$n \equiv$ slope of a straight line.

Data from 11 detonations during May, June and July of 1958 (Sh57) indicated a mean fallout deposition exponent of 18.8. This mean value was observed at Utirik, Rongelap, Parry and Wothe and applied to early time post detonation of

BRAVO to obtain the initial increasing exposure rate history evinced on figures 24 and 26. This method yielded a fallout deposition period of 5.5 hours on Rongelap and 12 hours on Utirik. This time compares well with the original observations reported by the Marshallese and by U.S. Navy personnel stationed in the area (SI 57). Initial dose equivalents on "acute doses" are developed in greater detail in another report.

Figure 25 demonstrates the external exposure following the 1958 testing series. Since return to Rongelap followed 3 years after the BRAVO contamination, this series contributed in large part to the external exposure post return.

SUMMARY

The Castle BRAVO shot of March 1954 caused the contamination of the inhabited atolls Rongelap and Utirik. Evacuation from Rongelap commenced 50 hours after detonation and from Utirik 55 hours after detonation. During June of 1954 and June of 1957 the return of the Utirikese and Rongelapese occurred respectively. Body burden data for dosimetrically significant nuclides was obtained throughout the residence interval post return primarily by direct in vivo gamma spectroscopy and by indirect bioassay techniques.

The dosimetric models used in this analysis were representative of a declining continuous uptake regime. Dietary decline of radioactivity included radioactive decay of the source and a conglomerate of other factors which might have included increased use of imported foods and weathering of the source. Dietary loss rate constants were estimated from sequential body burden data and were comparable for both atolls.

Variation in body burden history data for a particular nuclide on a particular atoll was observed in whole body counting data and urine bioassay results. This was attributed principally to the statistical variation encountered when small groups are sampled from a heterogeneous group of body burdens in people and in the case of urine bioassay, additional variation was introduced during the laboratory analysis of samples.

Daily activity ingestion rates were determined for all measured radionuclides. In general, infants, children and adults between 20-40 years of age ingested more activity each day than did adolescents and persons greater than 40 years of age. Maximum deviation from the average value of the daily activity ingestion rate for members of an age sub-group was no greater than a fac-

tor of 3. However, the population distributions illustrated a maximum factor of 5 times the mean activity ingestion rate value.

Dose equivalent rates post return were determined for members from both atolls. For Rongelap Atoll, the residents received approximately 100-200 mRem per year during the first 5000 days post return from internal emitters. The principle contributing nuclide was ^{137}Cs . For Utirik Atoll, the residents received 10-15 Rem per year during the first 400 days post return. The major contributing nuclides were ^{65}Zn and ^{60}Co . Dose equivalent rates to the Utirikese from internal emitters fell below 500 mRem per year at approximately 1200 days post return.

The dose equivalent for population sub-groups and for individuals was determined. Table 6 summarizes the results for the total body, thyroid, red marrow, testes, ovaries, lower large intestine wall and the liver. The catenary compartment model of Bernard and Hayes (Ber70) was used to determine doses to various segments of the gastrointestinal tract. The Utirikese received significantly more radiation dose from ^{65}Zn , ^{60}Co , and ^{55}Fe than did the Rongalapese due to the short mean residence times of these nuclides in the environment. ^{90}Sr doses to the Rongalapese were 2.5 times greater and ^{137}Cs doses 1.5 times greater than doses received by persons at Utirik. This occurred even though Utirik residents returned to their atoll 3 years earlier and it reflects the degree to which Utirik was less contaminated than Rongelap.

Table 6

Chronic Phase

Dose Equivalent Summary, Rem

Nuclide	<u>Total Body</u>		<u>Thyroid</u>	
	<u>Utirik Adults</u>	<u>Rongelap Adults</u>	<u>Utirik Adults</u>	<u>Rongelap Adults</u>
⁹⁰ Sr	.0118	.0267	.000749	.00169
⁵⁵ Fe	.0329	.0230	.0594	.0415
¹³⁷ Cs	1.13	1.71	1.55	2.35
⁶⁰ Co	.507	.0143	.359	.0101
⁶⁵ Zn	12.5	.0757	11.1	.0672
Total Internal	14.2	1.85	13.1	2.47
Net External	3.19	2.02	3.19	2.02
Total	17.4	3.87	16.3	4.49

Nuclide	<u>Red Marrow</u>	<u>Rongelap Adults</u>	<u>Testes-Ovaries</u>	
	<u>Utirik Adults</u>		<u>Utirik Adults</u>	<u>Rongelap Adults</u>
⁹⁰ Sr	.0537	.123	.000749-.000749	.00169-.00169
⁵⁵ Fe	.0603	.0422	.0583-.0620	.0736-.0433
¹³⁷ Cs	1.70	2.57	1.54-1.74	2.33-2.63
⁶⁰ Co	.629	.0177	.443-1.78	0.12-.0502
⁶⁵ Zn	17.2	.103	11.3-16.3	.0685-.0988
Net Internal	19.6	2.86	13.3-19.9	2.49-2.82
Net External	3.19	2.02	3.19	2.02
Total	22.8	4.88	16.5-23.1	4.51-4.84

Table 6 (Cont'd)

Chronic Phase

Dose Equivalent Summary, Rem

<u>Nuclide</u>	<u>Lower Large Intestine Wall</u>		<u>Liver</u>	
	<u>Utirik Adults</u>	<u>Rongelap Adults</u>	<u>Utirik Adults</u>	<u>Rongelap Adults</u>
⁹⁰ Sr	.225	.567	.000671	.00152
⁵⁵ Fe	.0666	.0465	.115	.0804
¹³⁷ Cs	.591	.895	1.81	2.74
⁶⁰ Co	4.66	.132	.792	.0223
⁶⁵ Zn	15.0	.0910	16.5	.136
Total Internal	20.5	1.73	19.2	2.98
Net External	3.19	2.02	3.19	2.02
Total	23.7	3.75	22.4	5.00

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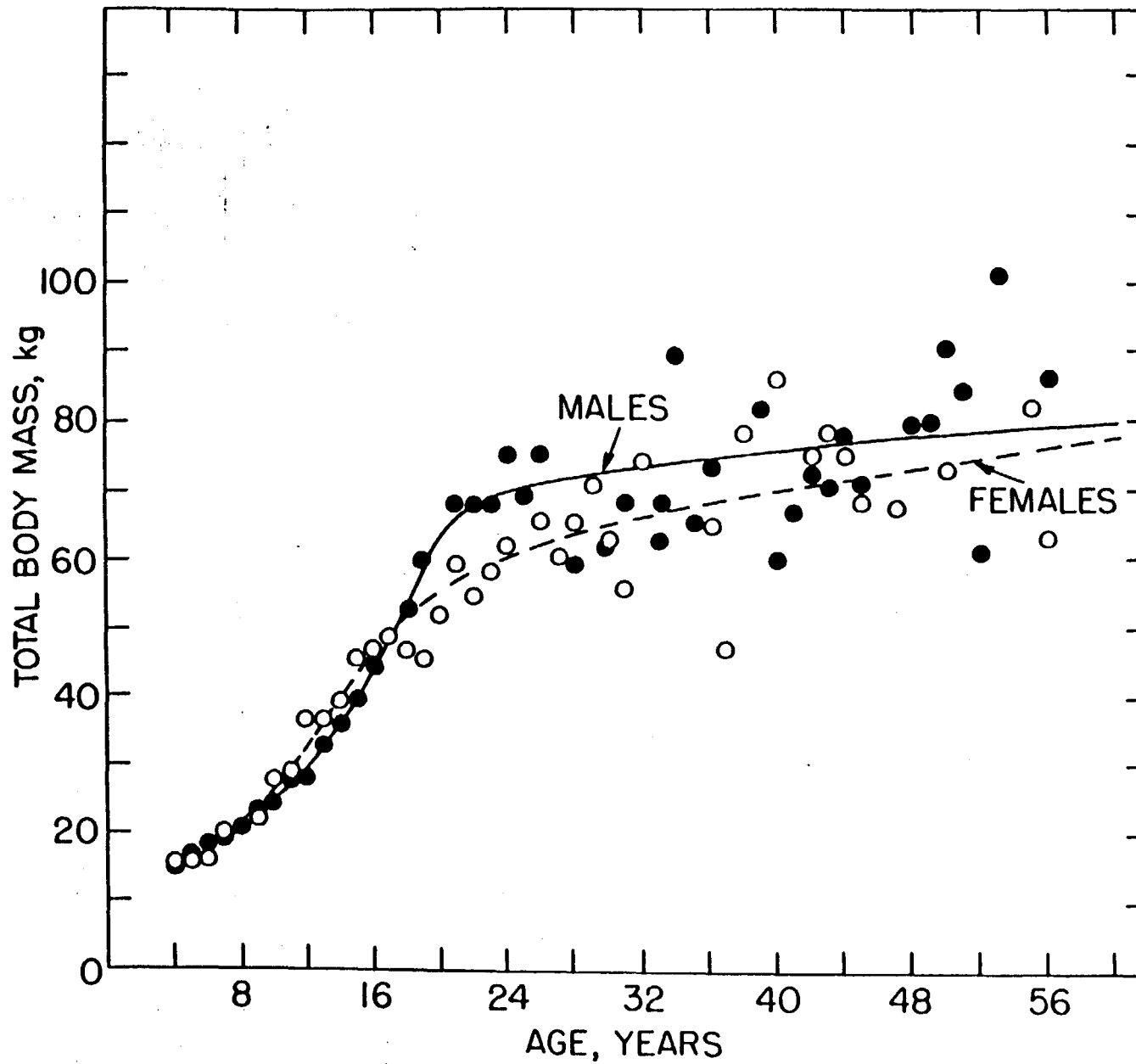


Fig. 1 Body Mass as a Function of Age for Residents of Rongelap Atoll

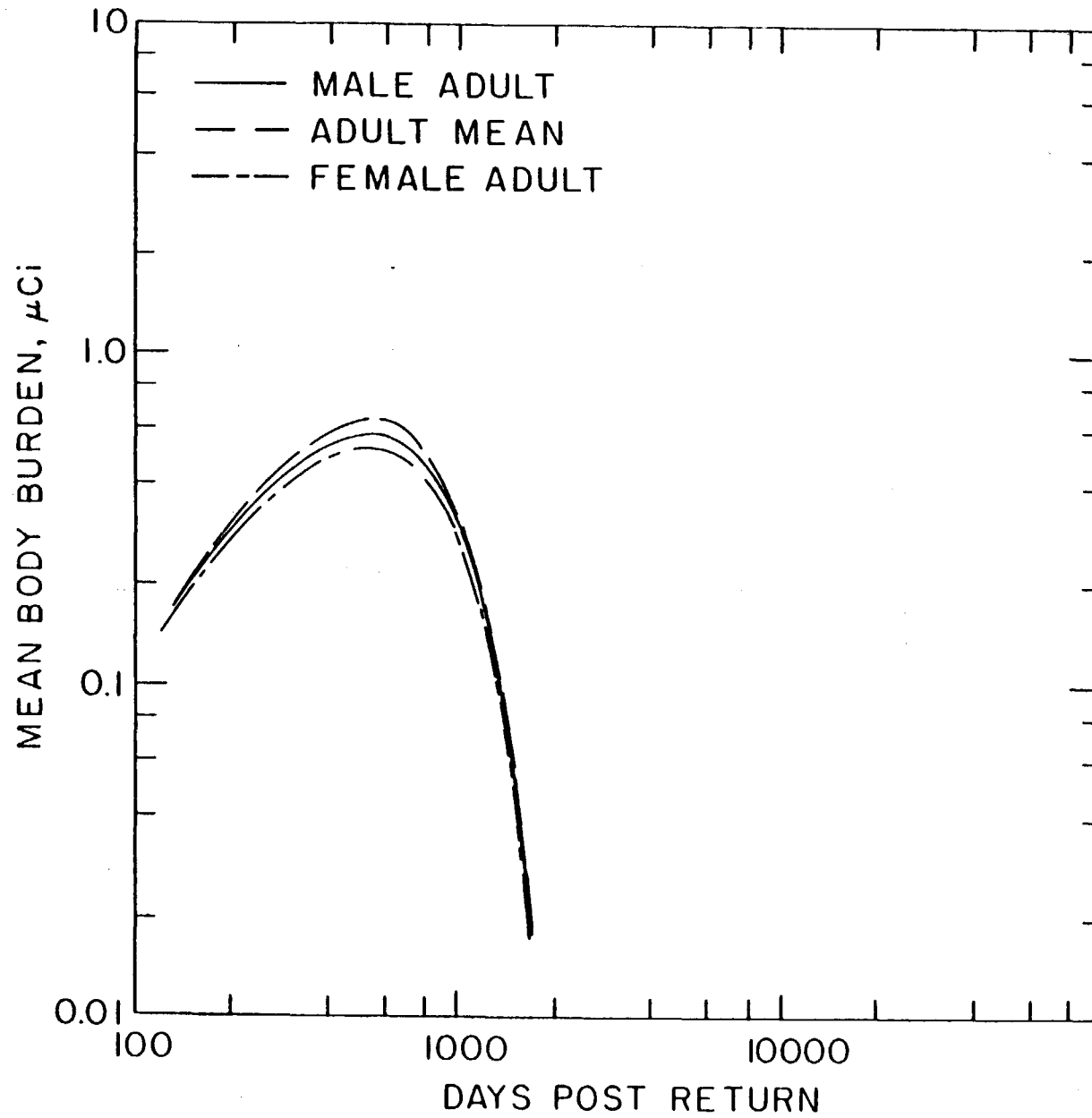


Fig. 3 Mean Adult ^{65}Zn Body Burden History
at Rongelap Atoll

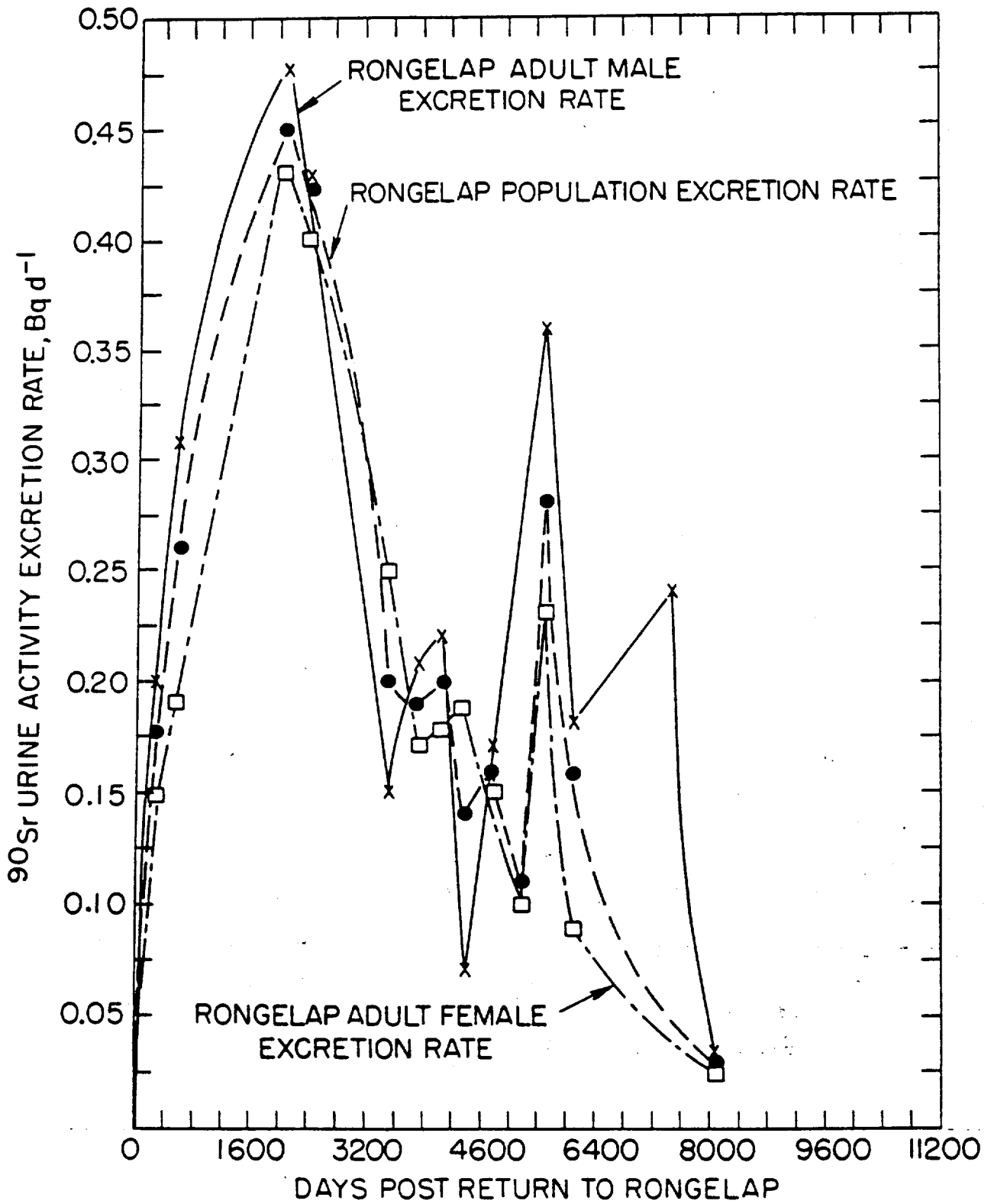


Fig. 5 Mean Adult ⁹⁰Sr Urine Activity Excretion Rate at Rongelap Atoll

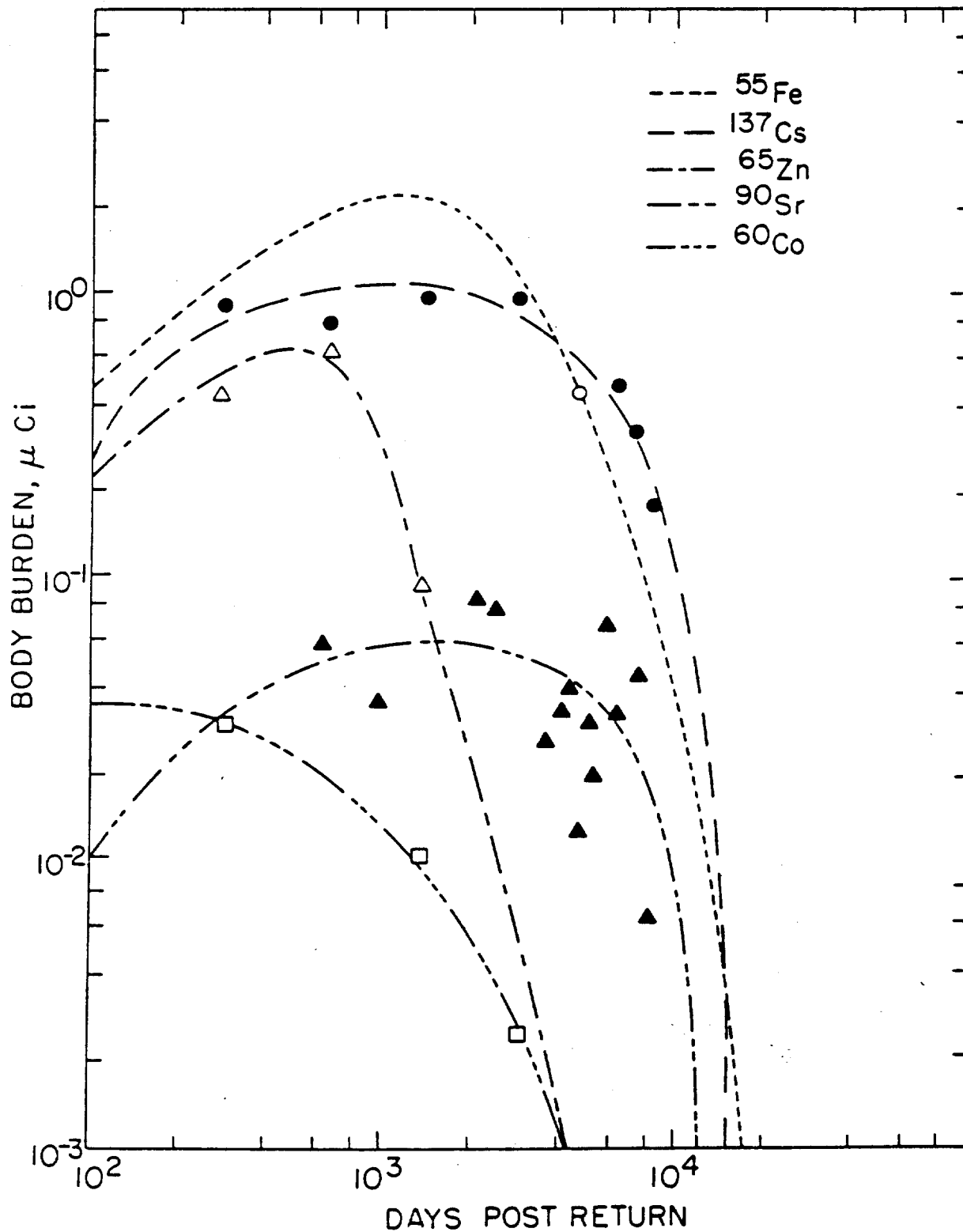


Fig. 7 Composite Nuclide Body Burden History
 for Adults at Rongelap Atoll

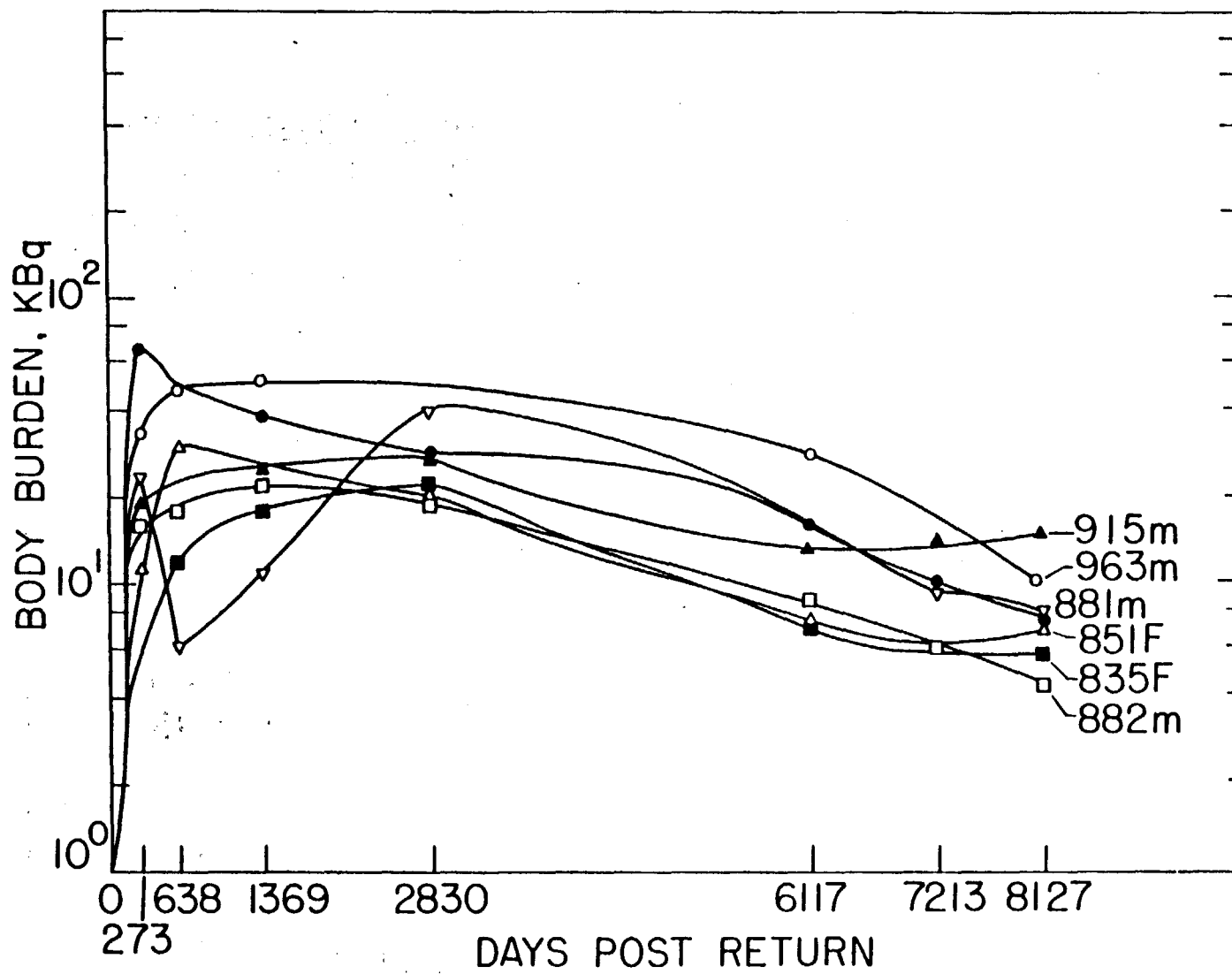


Fig. 9 Individual Male and Female Body Burden Histories Randomly Chosen from the Rongelap Atoll Population

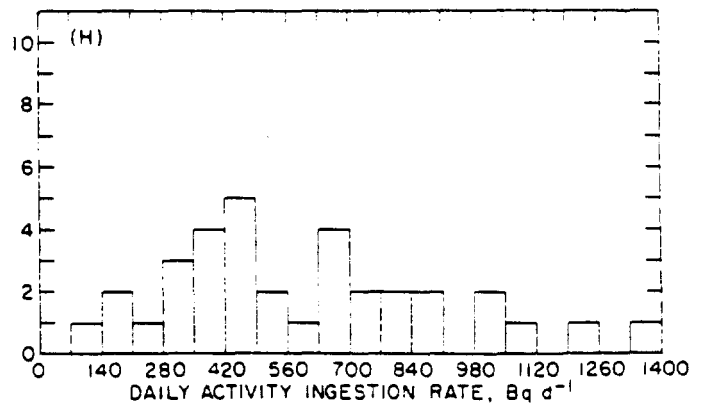
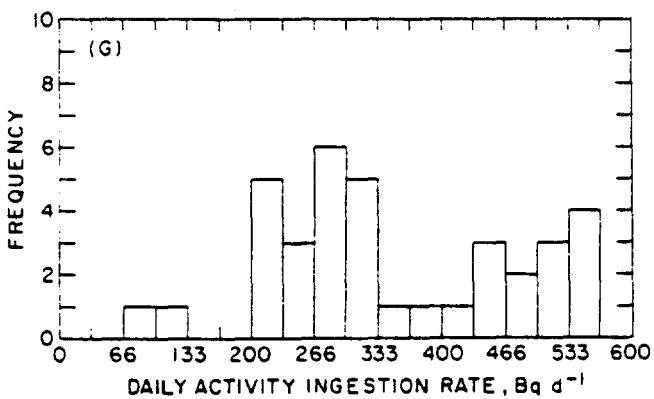
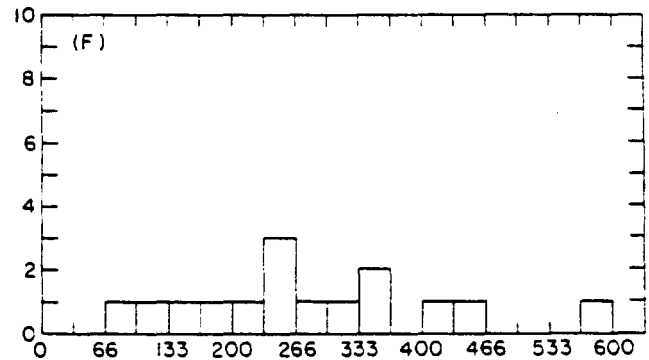
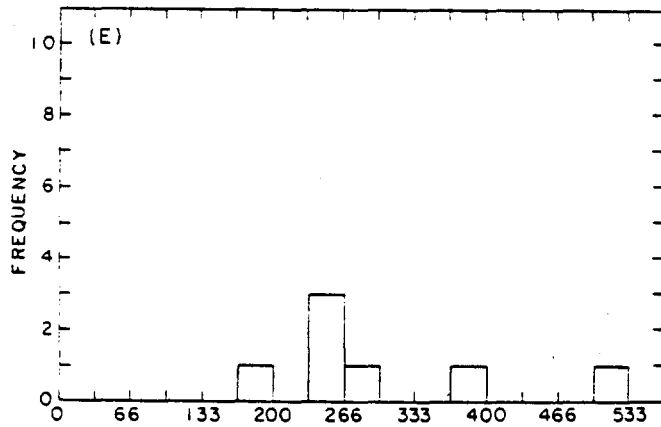
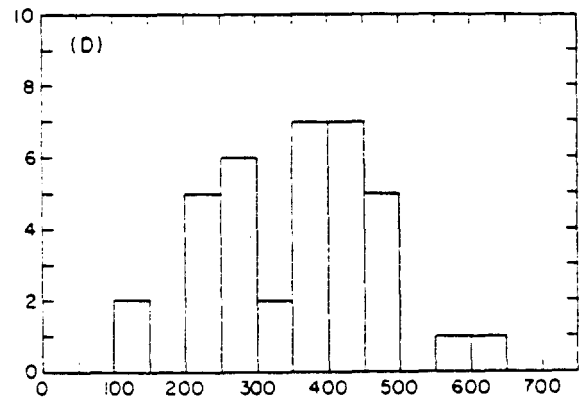
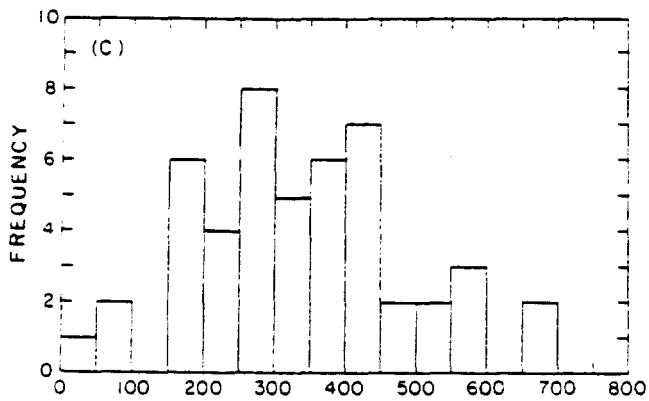
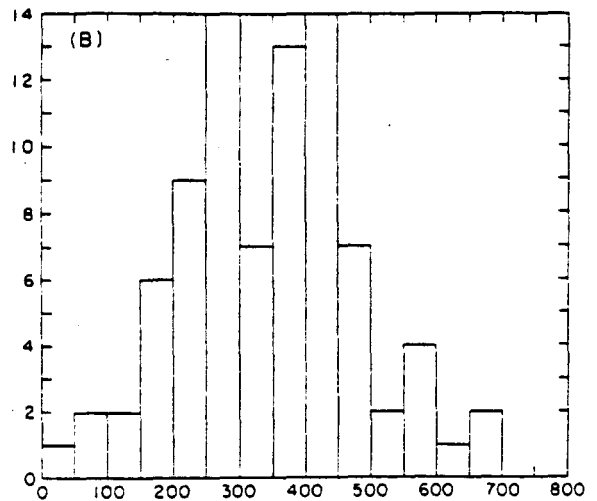
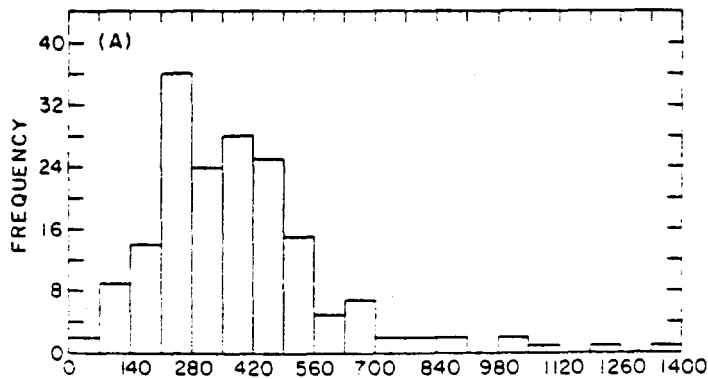


Fig. 11 ¹³⁷Cs Daily Activity Ingestion Rate for (A) All Residents (B) Adults (C) Adult Males (D) Adult Females (E) Young Adults (F) Adolescents (G) Children and (H) Infants

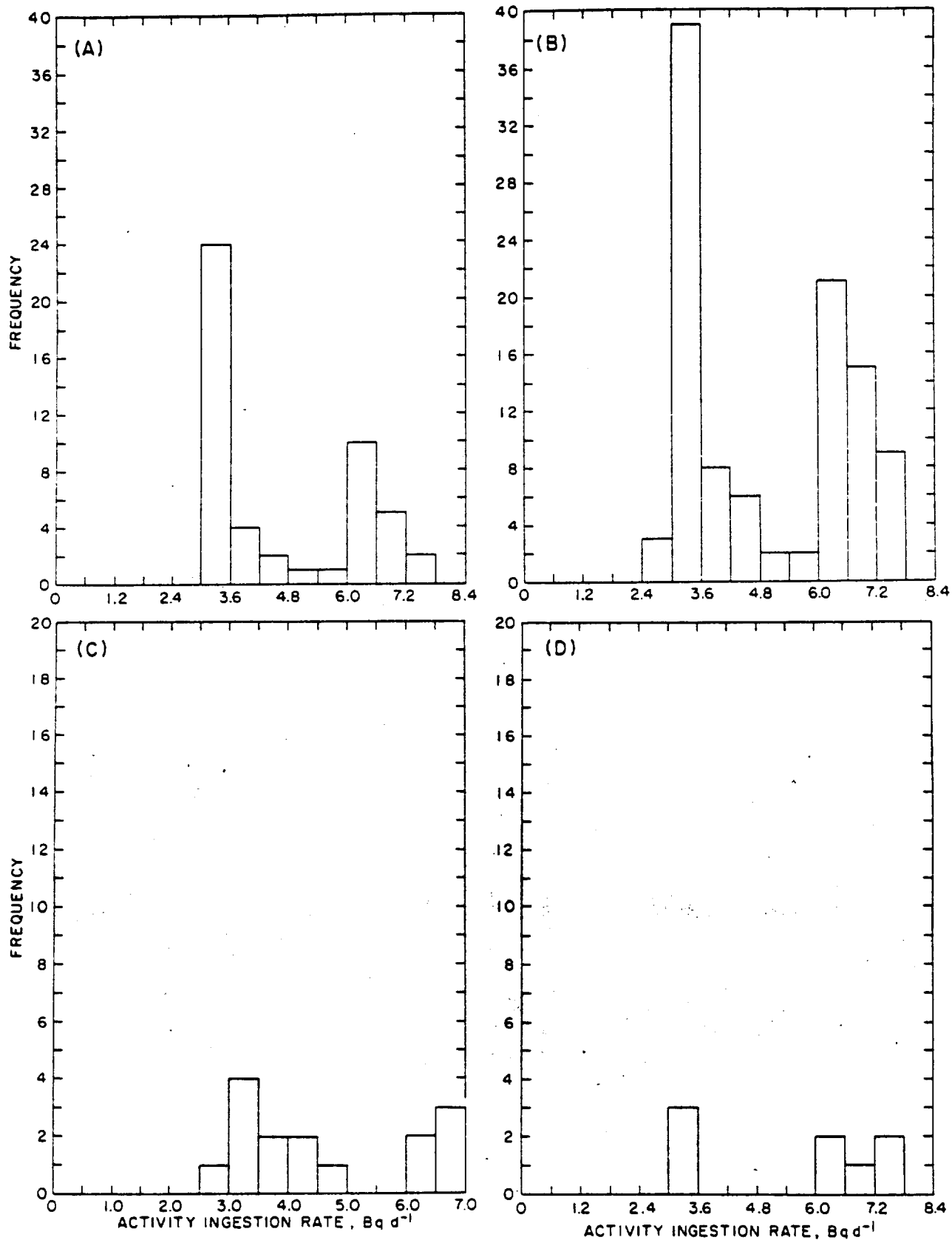


Fig. 13a ^{90}Sr Daily Activity
Ingestion Rate for (A) Adults
(B) All Residents (C) Infants
and (D) Adolescents on Rongelap

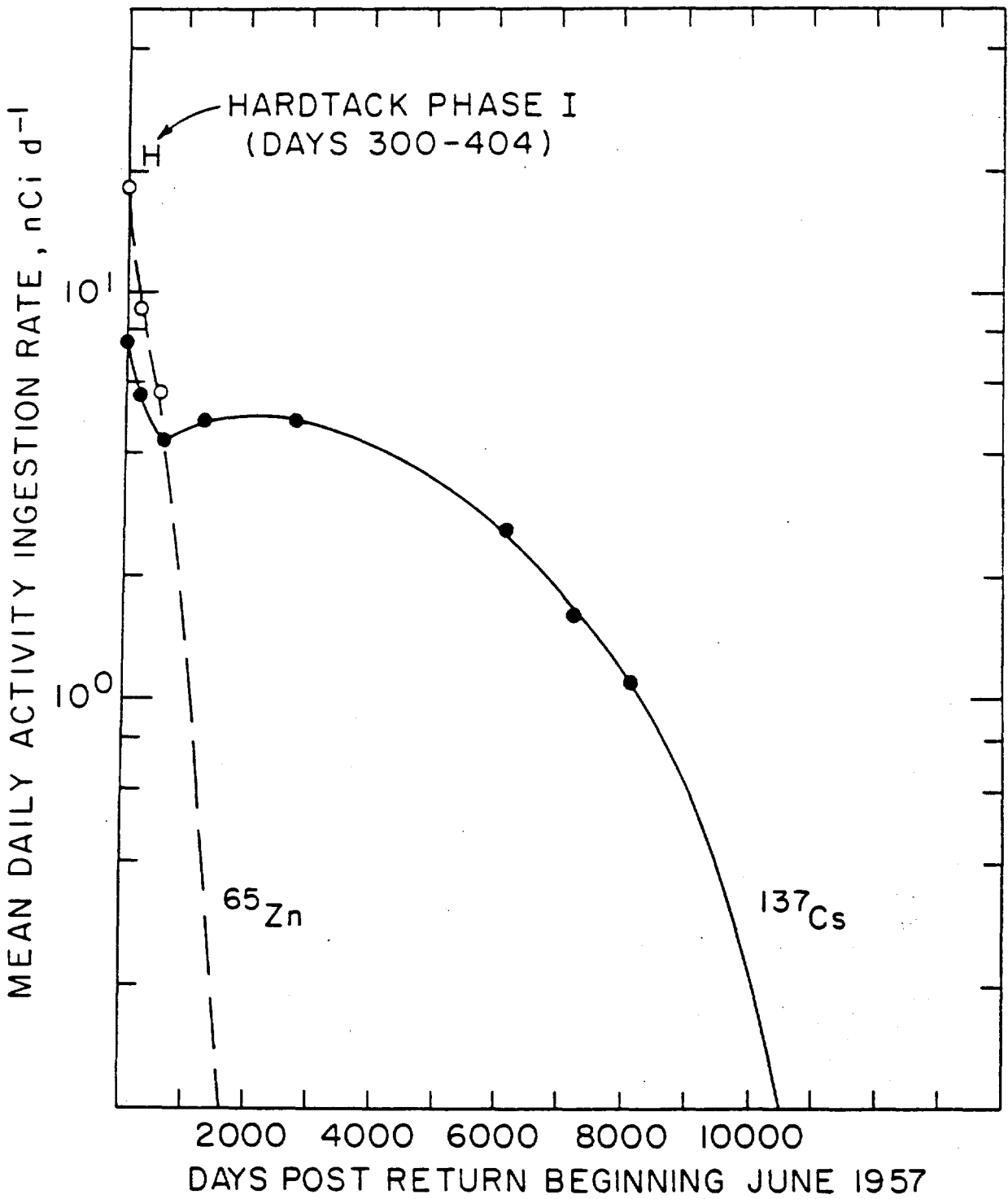


Fig. 14 Adult Mean Daily Activity Ingestion Rate for ¹³⁷Cs and ⁶⁵Zn at Rongelap Referenced to Mid-1957

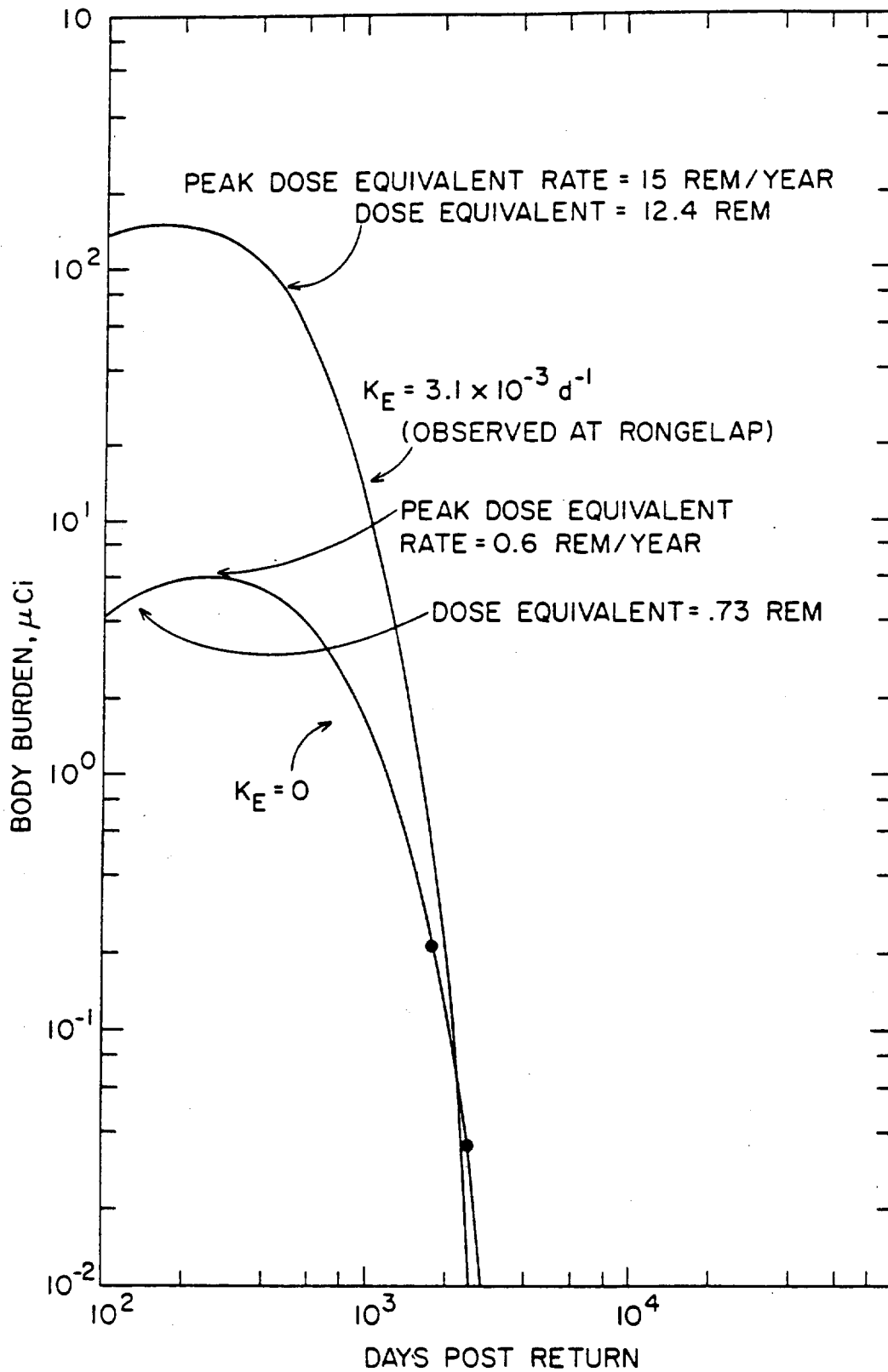


Fig. 16 Mean Adult ^{65}Zn Body Burden, Peak Dose Equivalent Rate and Dose Equivalent for Utirik Atoll

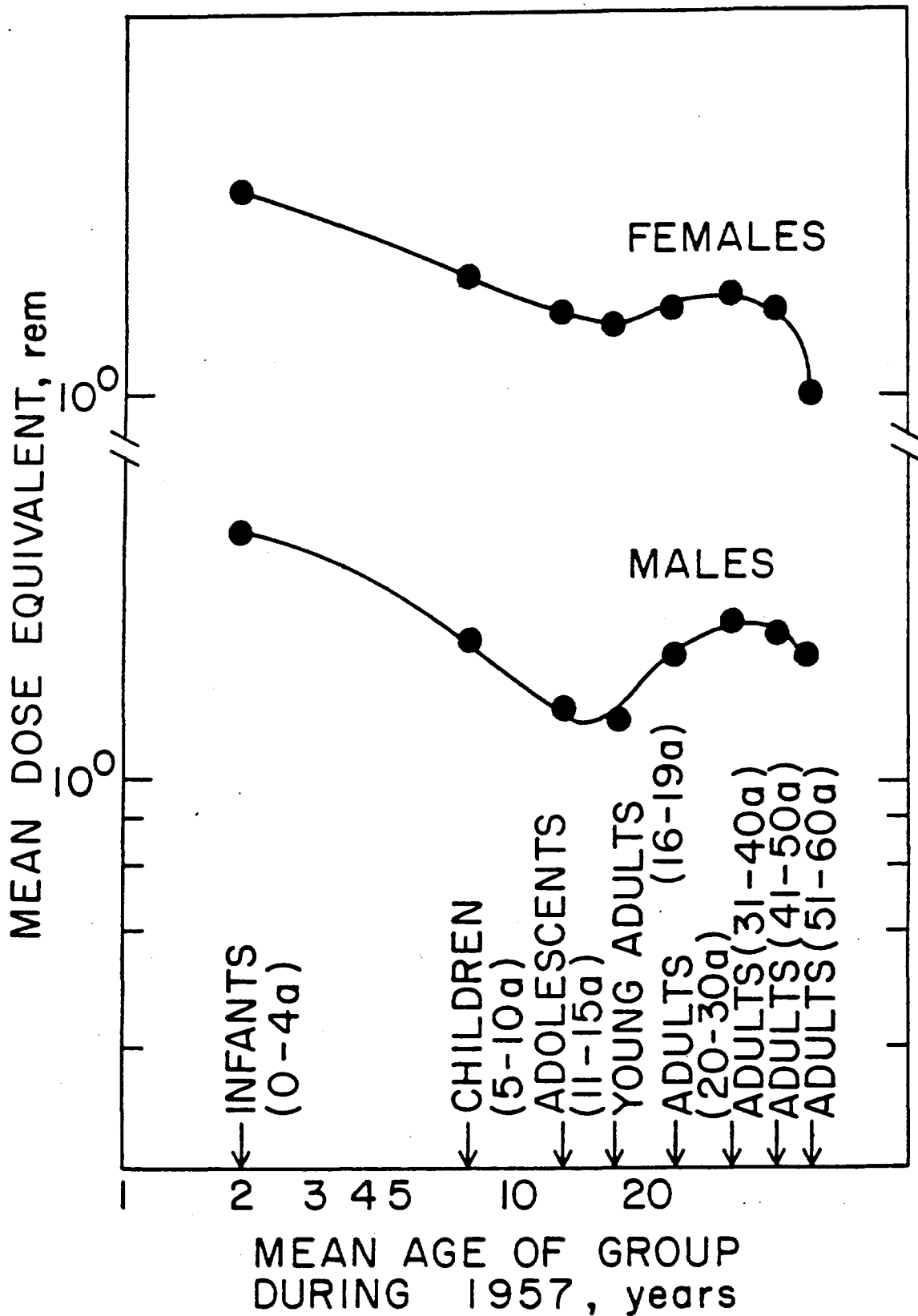


Fig. 18 ¹³⁷Cs Mean Dose Equivalent for Various Mid-1957 Age Groups for the Interval 1957 to 1980 at Rongelap Atoll

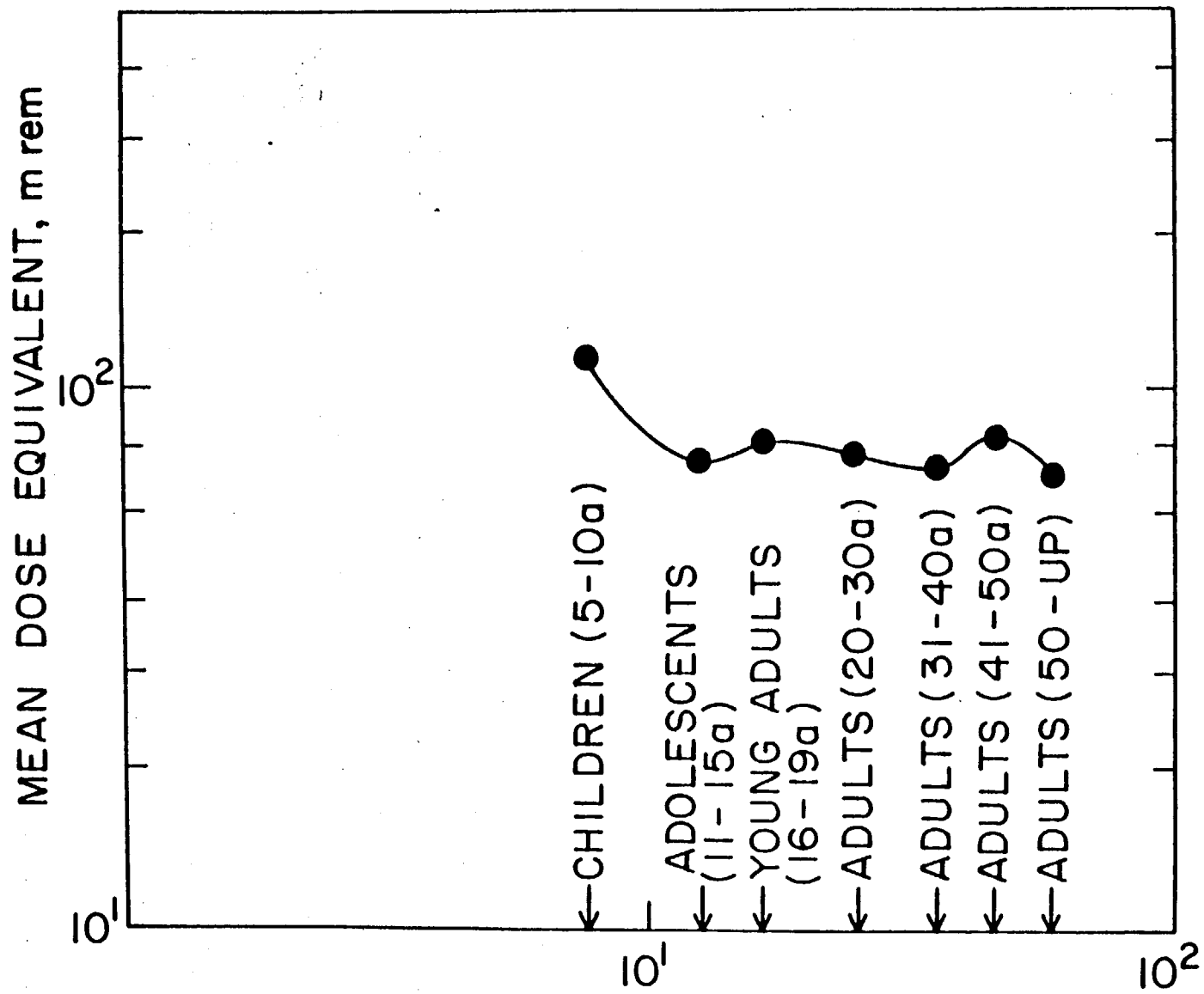


Fig. 20 ^{65}Zn Mean Dose Equivalent for Various Mid-1957 Age Groups for the Interval 1957 to 1980 at Rongelap Atoll

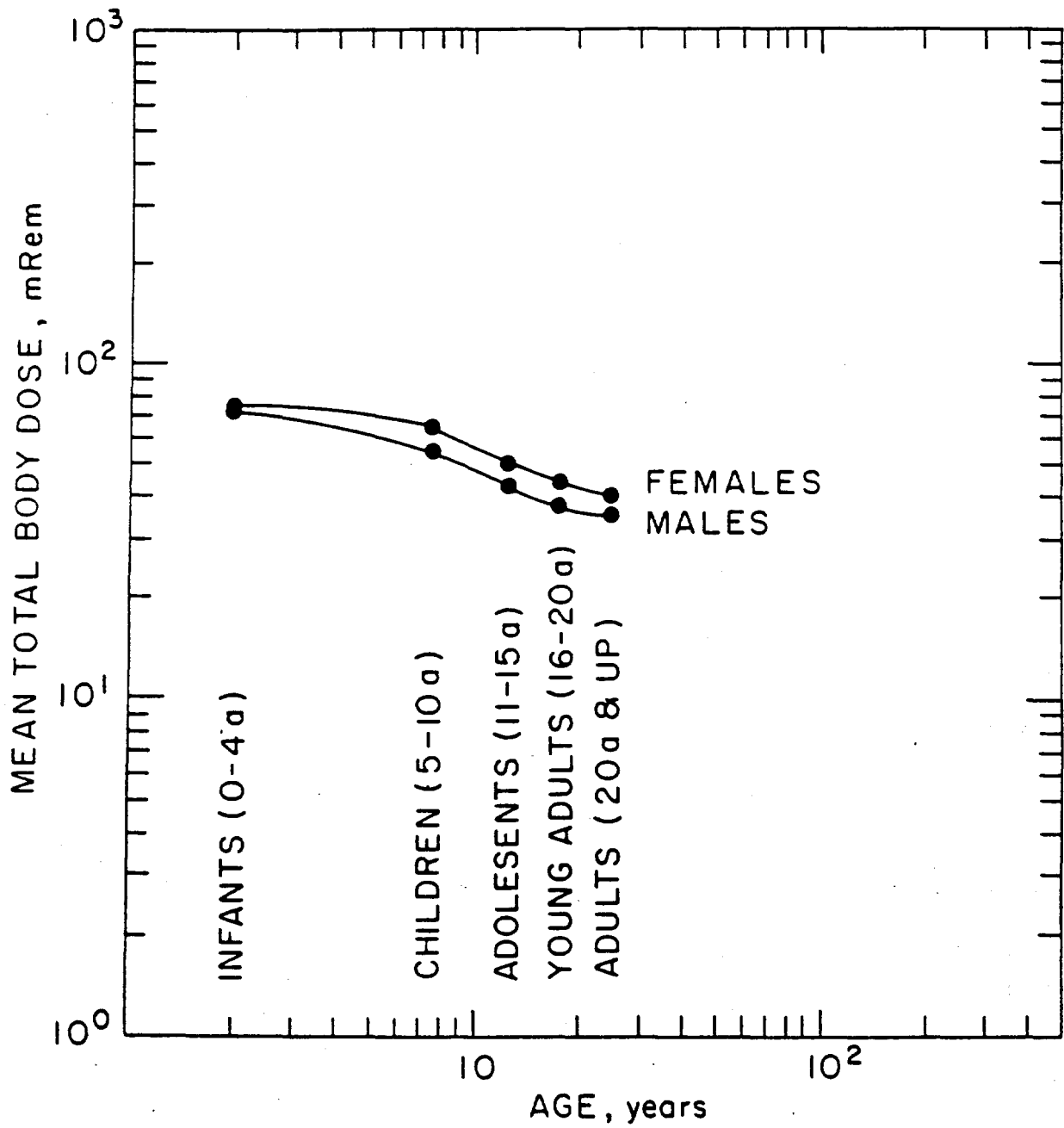


Fig. 22 Age and Sex Group
 Mean Values for ^{90}Sr
 Dose Equivalent for the
 Interval 1957 to 1980 at
 Rongelap Atoll

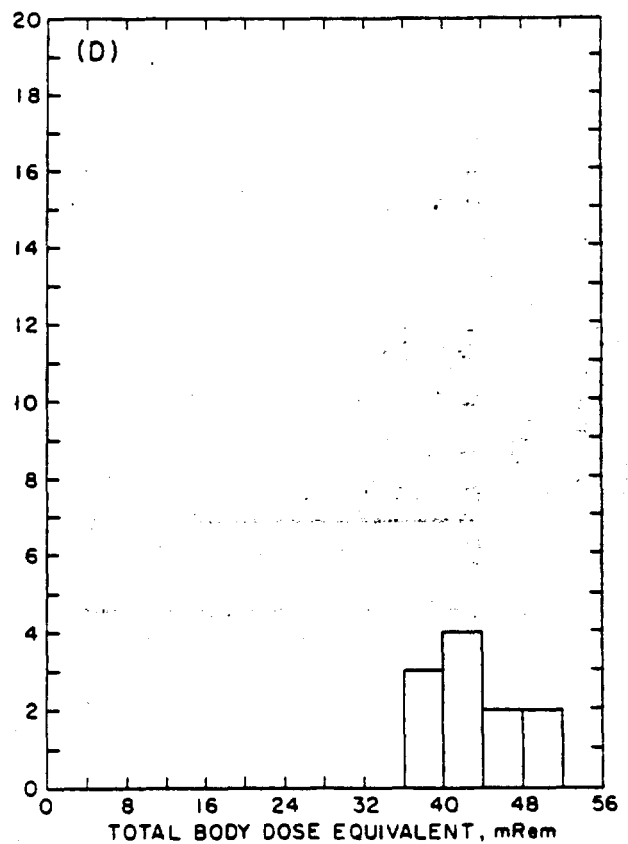
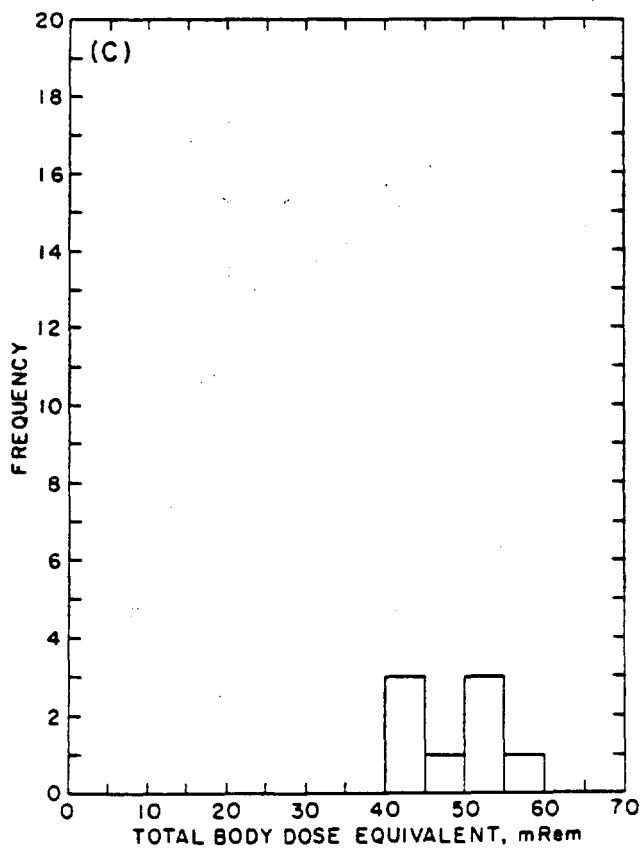
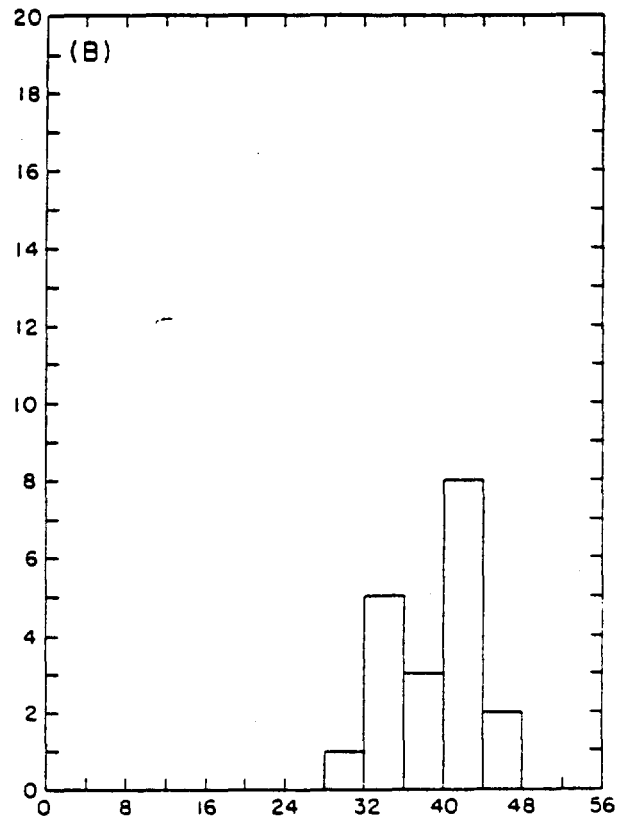
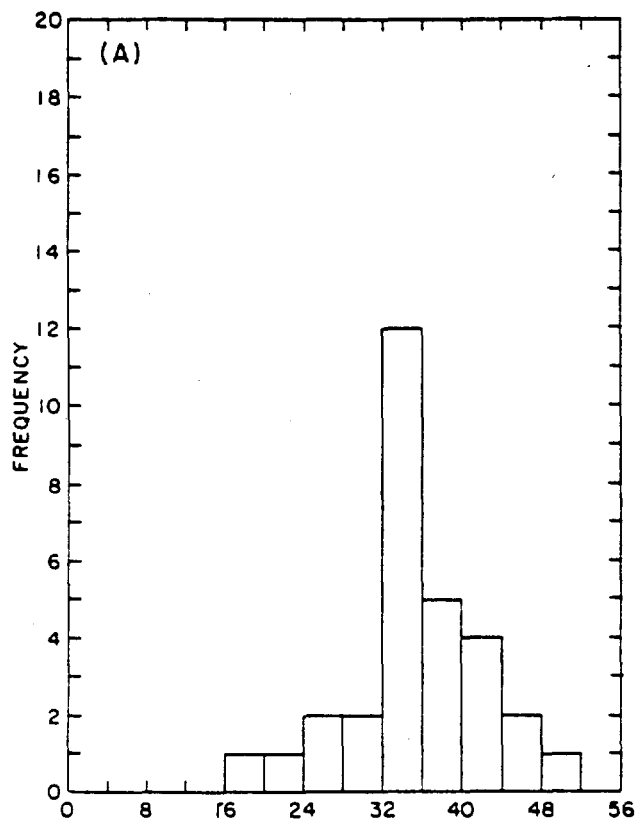


Fig. 23b ⁹⁰Dose Equivalent for
 (A) Adult Males (B) Adult Females
 (C) Adolescents and (D) Young Adults
 on Rongelap

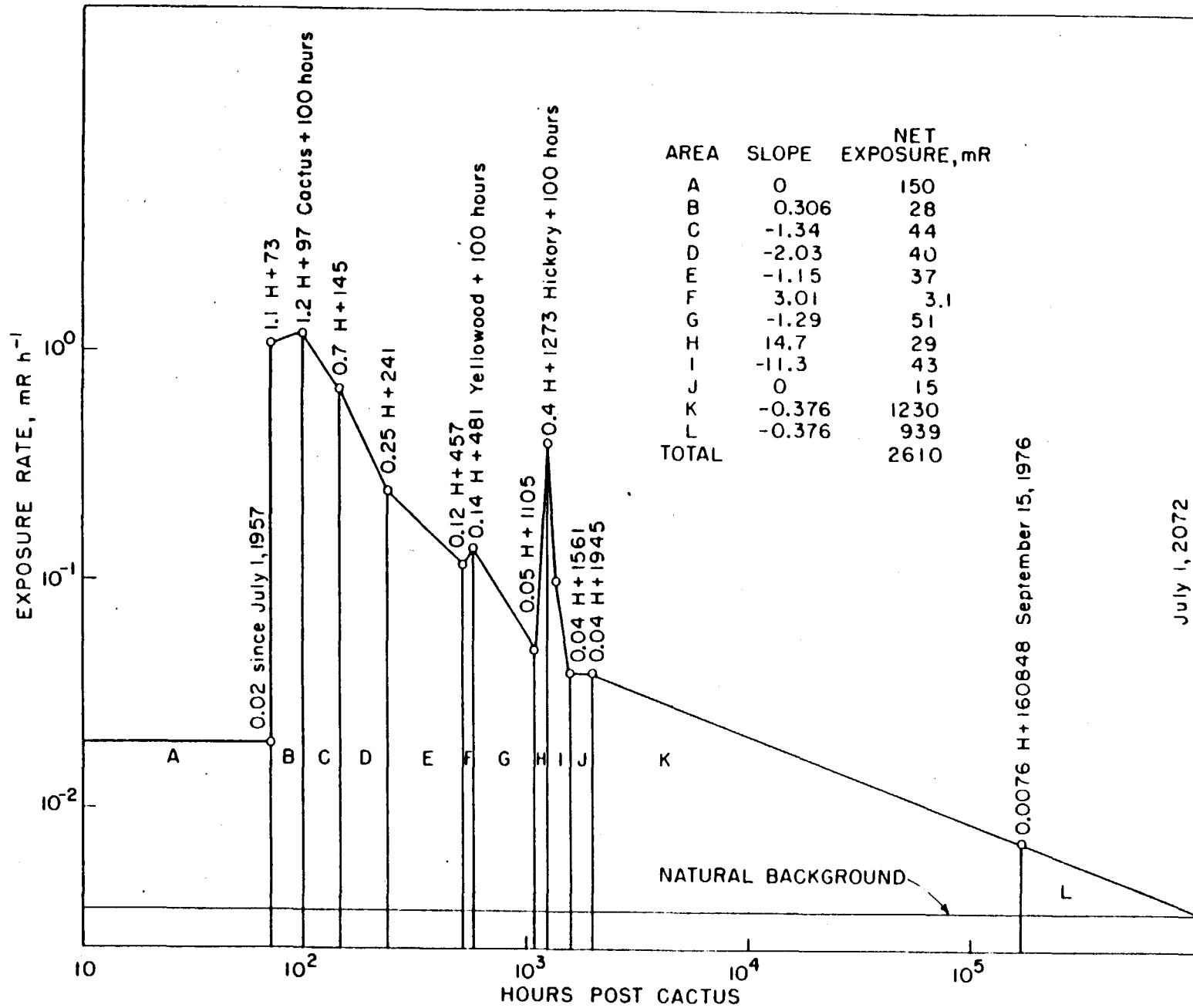


Fig. 25 Rongelap External Exposure Rate History Post Cactus